

Program Report 2000-P003

1998 Annual Status Report

Status and Trend of Submersed and Floating-leaved Aquatic Vegetation in Thirty-two Backwaters in Pools 4, 8, 13, and 26 and La Grange Pool of the Upper Mississippi River System



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Status and Trend of Submersed and Floating-leaved Aquatic Vegetation in Thirty-two Backwaters in Pools 4, 8, 13, and 26 and La Grange Pool of the Upper Mississippi River System

by

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June 2000

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Preface

The Long Term Resource Monitoring Program (LTRMP) was authorized under the Water Resources Development Act of 1986 (Public Law 99-662) as an element of the U.S. Army Corps of Engineers' Environmental Management Program. The LTRMP is being implemented by the Upper Midwest Environmental Sciences Center (UMESC), a U.S. Geological Survey science center, in cooperation with the five Upper Mississippi River System (UMRS) States of Illinois, Iowa, Minnesota, Missouri, and Wisconsin. The U.S. Army Corps of Engineers provides guidance and has overall Program responsibility. The mode of operation and respective roles of the agencies are outlined in a 1988 Memorandum of Agreement.

The UMRS encompasses the commercially navigable reaches of the Upper Mississippi River, as well as the Illinois River and navigable portions of the Kaskaskia, Black, St. Croix, and Minnesota Rivers. Congress has declared the UMRS to be both a nationally significant ecosystem and a nationally significant commercial navigation system. The mission of the LTRMP is to provide decision makers and river managers with information for maintaining the UMRS as a sustainable large river ecosystem given its multiple-use character. The long-term goals of the Program are to understand the system, determine resource trends and effects, develop management alternatives, manage information, and develop useful products.

This report presents the results of aquatic vegetation transect surveys conducted in 1998 by field station personnel under direction of the UMESC. Selected areas in Pools 4, 8, 13, and 26 of the Upper Mississippi River and La Grange Pool on the Illinois River were surveyed. This report satisfies, for 1998, Task 2.2.4.6, Evaluate and Summarize Annual Present-day Results under Goal 2, Monitor Resource Change of the Operating Plan (U.S. Fish and Wildlife Service 1993). The purpose of this report is to provide a summary of data regarding the distribution and abundance of submersed aquatic vegetation collected from the field stations for 1998. This report was developed with funding provided by the Long Term Resource Monitoring Program.

1998 Annual Status Report

Status and Trend of Submersed and Floating-leaved Aquatic Vegetation in Thirty-two Backwaters in Pools 4, 8, 13, and 26 and La Grange Pool of the Upper Mississippi River System

by

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Abstract: Thirty-two backwaters of the Upper Mississippi and Illinois Rivers were monitored for the eighth consecutive year in 1998 to determine the status and trend of changes of submersed and floating-leaved aquatic vegetation. Aquatic vegetation was sampled at regularly spaced sites along previously established permanent transects. Species compositions, frequencies of individual species, and the frequencies of sites that supported aquatic vegetation in 1998 were calculated and compared with results from previous years. The status and trend of aquatic vegetation in 1998 varied among the thirty-two backwaters. In upper Pool 4 (above Lake Pepin), where aquatic vegetation has declined since 1991, the trend continued in 1998. In lower Pool 4 (below Lake Pepin), where aquatic vegetation declined from 1991 to 1996 and started to rebound in 1997, the rebound continued in 1998. Aquatic vegetation has increased in Pools 8 and 13 since 1994 and is at or near its best (recorded) condition since 1991. In recent years, especially since a devastating flood in 1993 and through 1998, aquatic vegetation has been rare in La Grange Pool (Illinois River) and Pool 26. In Pool 26, however, aquatic vegetation was found in all four backwaters surveyed in 1998—three of the four had no aquatic vegetation in 1997. Similarly, aquatic vegetation was found in three of the four backwaters surveyed in La Grange Pool in 1998—one of the three had no aquatic vegetation in 1997.

Key words: Annual report, aquatic, floating-leaved, LTRMP, Mississippi River, submersed, vegetation

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Introduction

Aquatic vegetation in the Upper Mississippi River System (UMRS) has been monitored as part of the Long Term Resource Monitoring Program (LTRMP) since 1989. The 1998 growing season is the eighth consecutive year aquatic vegetation was monitored in selected backwaters for LTRMP. A report to summarize each year's data is a Program requirement, and this report satisfies the requirement for 1998 (U.S. Fish and Wildlife 1993).

The sampling scheme is referred to as transect sampling because sampling is conducted along permanent transects. The objective of the transect sampling is to reveal the status and trend of changes of the aquatic vegetation in the UMRS backwaters. This annual report provides basic data summaries, including the species composition of aquatic vegetation, species frequencies of presence, and the frequencies of presence of submersed or floating-leaved aquatic vegetation in individual backwaters.

Historical Perspective

Illinois River

A detailed description of the geological history and changes during the past 2 centuries of the Illinois River is in Bellrose et al. (1979) and Mills et al. (1966). The Illinois River has numerous backwater lakes, many of them expansive. Until the beginning of the 20th century, the backwaters were high-quality, near-pristine habitats for aquatic plants. A sequence of overlapping catastrophic events, however, drastically reduced and degraded the river's natural habitats.

Before 1900, a series of low-head dams were built at Marseilles, Henty, Copperous Creek, La Grange, and Kampsville. On January 1, 1900, completion of the Chicago Sanitary and Ship Canal allowed for diversion of water from Lake Michigan, washing sewage downstream and raising water levels along the Illinois by several feet. Subsequently, the river and its backwaters were severely polluted. Beginning in the 1900s, backwaters were isolated from the river with levees and dewatered for

agriculture. This, coupled with completion of the locks and dams from Dresden Heights to La Grange in the 1930s, added to the alteration of the river and its floodplain. Submersed aquatic plants began to disappear from the river as early as 1915 (Richardson 1921). Aquatic and marsh vegetation all but disappeared from the Illinois River during the period between 1920 and 1952.

Pollution from sedimentation presently is more troublesome than urban and industrial pollution because the former is cumulative. Storage capacity within the remaining bottomland lakes along the Illinois has been greatly reduced because of sedimentation. No new backwaters are being formed because of levees and other structures in and along the rivers. Many of the remaining backwaters are controlled to various degrees by levees, water gates, and water pumps. The hydrograph of the modern river is much more spiky and erratic than that of the natural river. Although water quality has improved over the last 20 years, sizable vegetation beds are still rare.

Upper Mississippi River

In contrast with the Illinois River, the Upper Mississippi River harbors more extensive submersed aquatic vegetation now than it did under natural (preconstruction) conditions. Before the 1930s, when the river was not impounded, most of the backwaters and marshes were flooded in spring and dried out in summer and fall. Water-level fluctuations and seasonal drying periods limited the development of aquatic plants in the backwaters (Green 1947). During the 1930s, a series of low-head dams were constructed to raise the water level during low discharges so a 9-foot-deep (minimum) navigation channel could be maintained. The impoundment of the river has created, and since maintained, additional backwater and marsh areas.

Aquatic plants initially prospered because water levels were more stable. Sustaining the quantity and quality of aquatic areas in the Upper Mississippi River, however, is a matter of concern for a number of reasons. Structures have altered the scouring and deposition dynamics of the natural river, which

will likely inhibit the formation of new backwater areas. Bathymetric diversity in the impounded areas is declining and commercial towboat and recreational watercraft traffic is predicted to increase. Consequently, water turbidity may be increased by amplified wind-fetch effect. The LTRMP was created in response to these and other concerns (Rasmussen and Wlosinski 1988). Identifying the trends of changes of the UMRS biological resources is one of the primary goals of the LTRMP (U.S. Fish and Wildlife Service 1993).

Study Areas

The LTRMP has six trend analysis areas, referred to herein by the navigation pool designations according to the U.S. Army Corps of Engineers lock and dam system. They are Pools 4, 8, 13, and 26 and the Open River on the Mississippi River and La Grange Pool on the Illinois River (Figure 1). Pool 26 also includes 12 miles of the lower Illinois River upstream of its confluence with the Mississippi River. The backwaters selected for aquatic vegetation monitoring are distributed in five of the six LTRMP trend analysis areas. The Open River LTRMP trend analysis area is not included in aquatic vegetation monitoring because it lacks any sizable backwaters or stable aquatic vegetation beds connected with the river.

Seven backwaters were surveyed in Pool 4 in 1998 (Figure 2). Dead Slough Lake, Goose Lake, Mud Lake, and Catherine Pass (referred to as Bay City Flats in reports of monitoring before 1995 [Rogers et al. 1998; Appendix A]) are upstream of Lake Pepin. (Lake Pepin is a large tributary delta lake created by the deposition of vast amounts of sand at the confluence of the Chippewa and Mississippi Rivers, near the center of Pool 4.) The other three backwaters, Robinson Lake, Peterson Lake, and Big Lake are below Lake Pepin. Big Lake consists of a complex of smaller backwaters where three clusters of transects were established, including one cluster in Rice Lake and another in Big Lake Bay. Because Rice Lake and Big Lake Bay are similar in habitat conditions and differ from the rest of Big Lake, and because sample sizes from the two areas are individually small, Rice Lake and Big Lake Bay are treated as one backwater

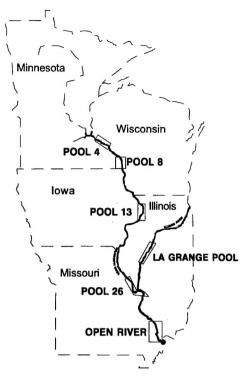


Figure 1. Location of Navigation Pools 4, 8, 13, and 26 and La Grange Pool in the Upper Mississippi River System where aquatic vegetation was surveyed, Long Term Resource Monitoring Program, 1998.

separate from the rest of Big Lake: Rice Lake-Big Lake Bay. The transects in Peterson Lake were also divided into Lower and Upper Peterson Lake groups and treated as two separate backwaters. All seven backwaters have been surveyed every year since 1991. In addition to the seven backwaters. Upper Mud Lake was surveyed from 1993 to 1996. Surveying of Upper Mud Lake was discontinued in 1997 because it is semi-isolated and atypical of backwaters in upper Pool 4 and access and sampling became excessively difficult. Consequently, the transects in Pool 4 were grouped into nine backwaters in analysis and reporting. A Habitat Rehabilitation and Enhancement Project (HREP), which involved dredging part of Big Lake Bay, was completed in spring 1993. Construction of an HREP in Peterson Lake in fall 1995 required relocating the first transect downstream about 30 m.

Eight backwaters were surveyed in Pool 8 in 1998 (Figure 3). Target Lake, Lawrence Lake, a backwater near Goose Island (herein referred to as the Goose Island backwater), and Shady Maple are all backwaters contiguous to the main channel

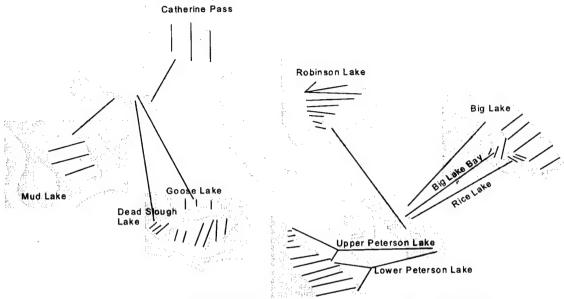


Figure 2. Location and arrangement of transects in Pool 4 (Upper Mississippi River) in 1998.

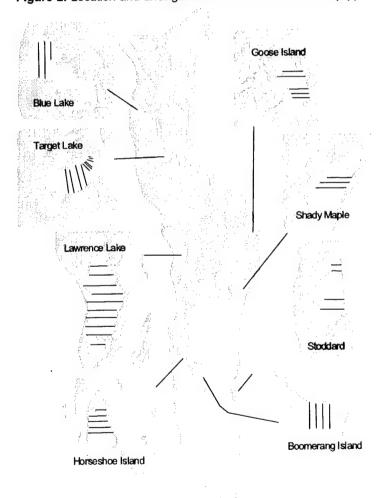


Figure 3. Location and arrangement of transects in Pool 8 (Upper Mississippi River) in 1998.

(Appendix A). Blue Lake and a backwater near Stoddard, Wisconsin (herein referred to as the Stoddard backwater), are isolated from the main channel. Transects established in the interior of Horseshoe and Boomerang Islands are in the impounded area of Pool 8. Target Lake, Lawrence Lake, Goose Island backwater, Shady Maple, and Horseshoe Island have been surveyed every year since 1991. The Stoddard backwater was added in 1992 and Blue Lake and Boomerang Island were added in 1993. Horseshoe Island and Boomerang Island are part of the Pool 8 Islands HREP.

Seven locations were surveyed in Pool 13 in 1998 (Figure 4). Brown's Lake, Savanna Bay, and Spring Lake are contiguous with the main channel, whereas Pomme de Terre, Potter's Marsh, Johnson Creek, and Johnson Creek Levee are in the impounded areas of Pool 13 (Appendix A). Brown's Lake and Potter's Marsh are HREP sites begun in 1988 and 1994, respectively. Transects in all seven locations were established in 1991. Both spring and summer sampling periods in 1998 were characterized by frequent

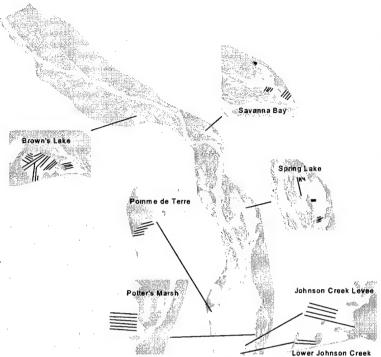


Figure 4. Location and arrangement of transects in Pool 13 (Upper Mississippi River) in 1998.

severe storms and mild temperatures. A 1-foot drawdown in the lower pool was begun 19 June, but was lost 3 July because of high flow conditions (flows greater than 110,000 cfs). It was restarted 11 July but was halted because of low flow conditions (flows less than 50,000 cfs).

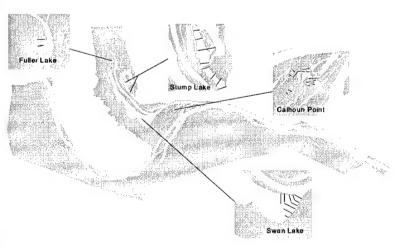
Four backwaters were surveyed in Pool 26, which includes portions of both the Mississippi and

lower Illinois Rivers (Figure 5). The four backwaters surveyed are in the lower 12 miles of the Illinois River. Fuller Lake, Calhoun Point, and Stump Lake are isolated from the flow of the Illinois River and are managed as moist soil units (mainly mimic preimpoundment conditions). The Calhoun Point area consists of several backwater lakes, sloughs, and ephemeral ponds. The fourth backwater, Swan Lake, is a large shallow area contiguous with the Illinois River. Although not presently managed as a moist soil

summer. Also, Swan Lake was undergoing HREP modifications, including a levee system that will convert this backwater to noncontiguous status. The levees and pumps will allow for moist soil unit management after 1998. Stump and Fuller Lakes were not sampled during summer 1998 because of low water levels. The Calhoun Point, Stump Lake, and Swan Lake transects were all established in 1991 (Appendix A). Fuller Lake was added in 1992.

Four backwaters were surveyed in La Grange Pool in 1998 (Figure 6; Appendix A). Banner Marsh and Spring Lake have been completely isolated from the flow of the Illinois River by agricultural levees for at least 5 decades. Point Lake is separated from the Illinois River by agricultural levees that have been overtopped frequently by flooding water. The

Grape Island location is a main channel border area and therefore directly affected by the river's flow. Banner Marsh was being modified as part of an HREP. Transects in Spring Lake, Point Lake, and Banner Marsh were established in 1991. Grape Island has been surveyed every year since 1992.



unit, Swan Lake is difficult to access Figure 5. Location and arrangement of transects in Pool 26 (Upper during low water levels in the Mississippi and Illinois Rivers) in 1998.

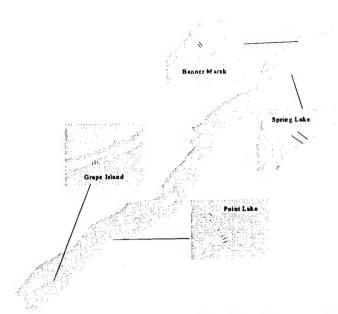


Figure 6. Location and arrangement of transects in La Grange Pool (Illinois River) in 1998.

Methods

Transect Sampling

The transect sampling scheme was designed to investigate aquatic vegetation along imaginary straight lines that traverse a backwater (or location), typically from shoreline to shoreline. These lines are traditionally called transects. The arrangement of transects in each backwater was determined in the first year (mostly during 1991) the backwater was chosen for monitoring (Figures 2–6). The transects are usually parallel lines of equal distance apart (50–200 m). During sampling, a field crew drives a boat along each transect and stops at regular distances (Appendix A). Each stop is called a sampling site.

Aquatic vegetation was sampled once each in spring and summer. Spring sampling is usually between May 15 and June 15 and summer sampling is usually between July 15 and August 31 (Rogers and Owens 1995). Since 1996, water temperature was utilized to determine the onset of spring sampling in addition to the rule of calendar dates. Accordingly, spring sampling begins when water temperature in the backwaters reaches 18° C, the temperature at which most submersed aquatic species have begun

to grow and elongate (Barko and Smart 1981; Madsen and Adams 1988; Flint and Madsen 1995). Sampling should be initiated by May 22 regardless of temperature so it can be completed in time. In 1998, spring sampling was between May 18 and June 23 and summer sampling was between July 21 and September 3 for all field stations (Appendix A).

The method of sampling aquatic vegetation at each site was modified from a technique used by Jessen and Lound (1962). The sampling area was a 2-m circle immediately in front of the bow of the boat. It was divided into thirds from which plants were collected using a long-handled thatching rake. The rake was

lowered to the bottom, twisted 180 degrees to snag vegetation, and then retrieved. The thatching rake had a 38-cm (15-inch) head with 20, 12.7-cm-long (5-inch) teeth. The area covered by the rake head during the twisting is approximately 0.1 m². Submersed plant species collected on the rake were identified and recorded. After all three segments of the circle were sampled, each species was assigned a rating of 1, 2, or 3 corresponding to the number of times the species was retrieved in the three rake samples. A rating of 4 was assigned if a species completely covered the rake teeth on all three twists. Beginning in 1997, a rating of "9" was given to species observed in the sampling area but not retrieved in any of the three rake samples. In previous years, such species were not recorded as legitimate entries. Water depth at each site was measured to the nearest decimeter using the handle of the rake, which is scaled and marked. Residual sediment retrieved on the rake during vegetation sampling was classified into one of three categories based on visual and tactile characteristics. These categories include silt and/or clay, mostly silt with some sand, and mostly sand with some silt. Sediment types were recorded for individual transects rather than for individual sites.

If a rooted floating-leaved species was present, it was assigned a rating of 1 to 4 based on the

amount of vegetative cover visible in the entire 2 m sampling area (1 = 1-25% cover, 2 = 26-50% cover, 3 = 51-75% cover, and 4 = 76-100% cover). A nonrooted floating species (e.g., Lemnaceae) was recorded only if it exceeded 5% of the surface area, but was excluded from analysis.

Fassett (1957), Voss (1972, 1985) and Gleason and Cronquist (1991) were the primary keys used for plant identification. Scientific nomenclature and common names are based on those found in the U.S. Department of Agriculture **PLANTS Database** on the Internet (http://plants.usda.gov/plants/). Leafy pondweed (Potamogeton foliosus) and small pondweed (P. pusillus) were usually not identified to species during field sampling and were recorded in a code "NLPW" for "small or leafy pondweeds." In some instances they were identified to species, then the record with the higher rating at the site was kept and renamed NLPW in analysis. Eurasian watermilfoil (Myriophyllum spicatum) includes specimens that were identified to species as well as specimens recorded as Myriophyllum spp.

Since 1991, voucher specimens have been collected and stored at each field station. Rare species and unusual specimens were sent to outside experts for verification.

Unusual Situation

Because of human error, species rating in Pool 13 in 1998 was assigned on a scale of 3 rather than 4. A submersed aquatic species was assigned a rating of 1, 2, or 3 corresponding to the number of times the species was retrieved in the three rake samples. Rating of a rooted floating-leaved species was assigned as follows: 1 = 1-33% cover, 2 = 34-67% cover, and 3 = 68-100% cover.

Data Summary

The frequency of a species is the number of sites in which the species were found, divided by the total number of the sites surveyed in the backwater, multiplied by 100 (percent). The frequencies of submersed or floating-leaved vegetation were similarly calculated. The frequency

of a species in any year was the higher value of the spring and summer frequencies. This treatment was based on considerations of the seasonality and the opportunistic nature of aquatic plant species in a river environment. The frequency data of 1998 were compared with frequency data of 1997 to determine whether an increase or decrease was statistically significant. Based on the assumption that the frequency of species in a backwater in any year is an independent variable that has a binomial distribution, a 95% confidence interval for the true difference between 1997 and 1998 was computed according to the following formula (Walpole and Myers 1978):

$$(p_1 - p_2) - 1.96 \sqrt{\frac{p_1 q_1}{n_1} + \frac{p_2 q_2}{n_2}} < (P_1 - P_2) < (p_1 - p_2) + 1.96 \sqrt{\frac{p_1 q_1}{n_1} + \frac{p_2 q_2}{n_2}}$$

where p_1 and p_2 are the frequencies of presence of a species in 1997 and 1998, respectively; n, and n_2 are the total number of sites sampled of 1997 and 1998, respectively; P_1 and P_2 are the unknown "true" frequencies of presence of the species in 1997 and 1998, respectively; and $q_1 = 100 - p_1$ and $q_2 = 100 - p_2$. If the interval contains the value zero, then the frequency of 1998 is concluded to be statistically the same as the frequency of 1997. Otherwise, we noted whether the 1998 frequency was an increase or decrease from the previous year. Also, we noted when a species was found for the first time in a backwater, when a species was present the previous year and not found in 1998, and when a species reappeared after a disappearance of one or more years. The frequency of a species is compared with available data for the period 1991–97 and noted whether it is the high or low for that period. Details are given for the individual pools in Results and Discussion following.

Results and Discussion

Pool 4

Based on data from nine backwaters, we concluded that the status and trend of aquatic vegetation in Pool 4 in 1998 differ between the upper pool (above Lake Pepin) and lower pool

(below Lake Pepin). Aquatic vegetation in the lower pool had another moderately good growing season in 1998 (Figure 7, Table 1). Compared with 1997, none of the five backwaters in the lower pool had decreased frequency of vegetated sites, whereas the frequency of submersed or rooted floatingleaved vegetation increased in Big Lake and the frequency of submersed vegetation sites increased in Robinson Lake. The 1998 growing season in the lower pool seemed to be an extension of a rebounding momentum that began in 1997. In contrast, aquatic vegetation in the upper pool had a poor growing season (Figure 7; Table 1). Compared with 1997, the frequency of submersed or rooted floatingleaved vegetation decreased in three-Goose, Dead Slough, and Mud Lakes-of the four backwaters surveyed. The 1998 growing season in the upper pool seemed to be an extension of a decline that began in 1992. Upper Pool 4 was species poor, typical since 1991; sago pondweed (Potamogeton pectinatus) was the only species found in 1998. Lower Pool 4, on the other hand, was species rich with 12 species of submersed and 2 of rooted floating-leaved plants found in 1998.

Wild celery (Vallisneria americana) was the most frequently encountered (MFE) species in Lower (19%) and Upper Peterson Lakes (27%) and Big Lake (14%). Curly pondweed (Potamogeton crispus) increased significantly in 1998 in Lower and Upper Peterson Lake. Eurasian watermilfoil was once again the MFE submersed species in Rice Lake-Big Lake Bay (22%), although it showed a nonsignificant decline in frequency from 1997. American lotus (Nelumbo lutea) remained the MFE aquatic species in these backwaters (31%). Water stargrass (Heteranthera dubia), the MFE species in Robinson Lake (43%), exhibited a significant increase in four of five lower pool backwater areas, including Upper Peterson Lake, Big Lake, Rice Lake-Big Lake Bay and Robinson Lake in 1998. Curly pondweed, flatstem pondweed (P. zosteriformis), Eurasian watermilfoil, Canadian waterweed (Elodea canadensis), and wild celery also showed significant increases in 1998 in Robinson Lake, the only backwater area in Pool 4 that had aquatic vegetation frequencies greater than 60%. Wild celery and water stargrass were both sampled at a high for the period 1991-98 along the

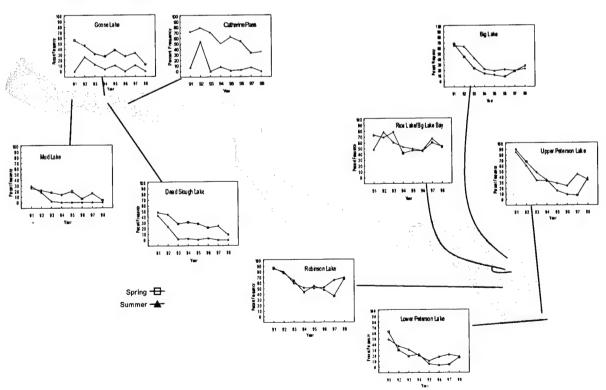


Figure 7. Frequency of aquatic vegetation in nine backwaters in Pool 4 (Upper Mississippi River), 1991-1998.

Table 1. Percent frequency of aquatic plant species by backwater or location in Pool 4, Upper Mississippi River, during the spring (Sp. May 26–June 17) and summer (Su, August 12–September 2) sampling periods, 1998.

		Lower	101	Upper	ě	Big		Rice Lake-	ake	Robinson	nosi	Catherine	ine	Goose		Dead Slough		Mud ake
		Peterso	eterson Lake Peterson Lake	Peterso	n Lake	La		Big Lake Бау	e Day	Lake	9	S C		2		Tion of		Ü
		Sp	Su	. ds	Su	Sp	Su	Sp	Su.	Sp	70	ď	20					- 1
Species	c	130	98	69	16	157	148	11	19	203	218	71	74	25 24	4 134	133	57	54
Submersed species																		•
	(Ceratophyllum demersum)	1.5	1.2	8.7	9.3		۵,	9.6	16.4	18.7	24.4						•	۰,
power p	(Zannichallia naluetrie)	,	,	4.36	1			1		1.0						•	•	•
	(Potomogaton crispus)	8		17.4	3.1	5.7	0.7	2.8	3.0€	27.64	4.	,				•	•	•
Ę	(P zosteriformis)	•		•	•	•			1.5	1.5	5.2 ^d		,		•		'	•
	(F. zosterye)		•		•		0.7	ŧ	3.0	0.5	0.5			•			'	•
pondweed, ionglear	(P. nectinatus)	6.2	1.2	8.8	5.2	2.5	•	•	3.0°	21.2	6.6	35.2	•	12.04	. 9.(- 9.0.6	3.54	an an
Il or leafy	(P. foliosus, P. pusillus)	•		4.3	2.1	9.0	•	2.8	0.9	3.9	6:1					•		•
water stargrass	(Heteranthera dubia)	•	3.5		17.54	3.8	7.24		10.4c.d.f	15.8	42.7c.d				·	•	•	•
urasian	(M. spicatum)	1.5	1.2	7.2	7.2	2.5	13	6.91	22.4	9.4	15.54			•			•	•
	(M. sibiricum)	•	•	•	,	,	•		۵,		ě,	,				•	•	•
watemymph, nodding	(Najas flexilis)	•	•		۵,	,	0.7		1.5		0.5	•		,		•	•	
waterweed, Canadian	(Elodea canadensis)	2.3	2.3	10.1	9.3	1.3	•	4.2	0.6	25.64	20.7	•			•			•
wild celery	(Vallisneria americana)	8.0	18.6	4.3	26.8	9.6	13.7	2.8	14.9°	25.1	42.3 ^{c.4}	•				•	•	
Rooted floating-leaved species	Sa																	
lotus, American	(Nelumbo lutea)		,		•	11.5	13.1°	31.0	17.9	1.5°	6.0				,	•		
waterlily, American white (Nymphaea odorata)	(Nymphaea odorata)	8.0	•		2.1	1.9	1.3	23.9	16.4	18.7	20.2	•				•		
Algae																		
filamentous algae	(Chlorophycaea)	•	•	1.4h	6.2 ^d	•	0.7h	4.2 ^{d.f}	•	5.9cd.f		•						
Frequency of submersed vegetated sites	etated sites	16.9	18.6	37.7	35.1	18.5	17.0	25.4	47.8	63.1	65.34	35.2		12.0°4	. 9	9.0.6	m	3.5%
Frequency of rooted floating-leaved vegetated sites	-leaved vegetated sites	8.0	٠	•	2.1	12.7	14.4	40.8	31.3	19.7	20.2	,						
Frequency of submersed and/or rooted floating-leaved sites	or rooted floating-leaved	16.9	18.6	37.7	35.1	27.44	22.9	53.5	52.2	67.0	69.5	35.2		12.068	. 9	9.0.6	. 3	3.5°6
*Located upstream of Lake Pepin. *Disappeared between 1997 and 1998. *Flighest record of 1991–1998.	epin. and 1998. 8.		Significa Lowest 1 Reappea	Significant increase from 1997. Lowest record of 1991–1998. Reappeared after absence in 1997.	se from 1 1991–19 absence i	997. 98. n 1997.	}		Sign hFirst	Significant	^k Significant decrease from 1997 ^h First record.	from 19	97.					

transects of Robinson Lake and Rice Lake-Big Lake Bay in 1998.

Sago pondweed reached a low in frequency for the period 1991–98 in three of four upper pool backwaters. Only one of four upper pool backwaters had aquatic vegetation frequencies greater than 15% during the spring sampling period. No aquatic plants were sampled in these backwaters during the summer sampling period.

Overall, 12 submersed and 2 rooted floating-leaved aquatic species were sampled in the 9 backwater areas of Pool 4 in 1998. Northern watermilfoil (Myriophyllum sibiricum), a native milfoil found in two backwaters in 1997, disappeared in 1998. The only species common to all nine backwaters was sago pondweed. Horned pondweed (Zannichellia palustris) was unique to spring sampling and nodding waternymph (Najas flexilis) to summer sampling. Robinson Lake once again had the highest species richness with all 14 species present.

Pool 8

According to data from eight backwaters, Pool 8 aquatic vegetation grew very well in 1998 (Figure 8, Table 2). The frequency of vegetated sites (submersed or rooted floatingleaved species) was greater than 70% in each of the eight backwaters. Horseshoe, Shady Maple, Lawrence Lake, and the Goose Island backwater had the highest frequencies since 1991. Shady Maple and Lawrence Lake increased from 1997, whereas Target Lake decreased. The frequency of vegetated sites was 100% at Blue Lake and the Goose Island backwater. Submersed vegetation increased in four backwaters and decreased in one, whereas rooted floating-leaved vegetation remained constant in all eight backwaters. A multiple number of species had increased or highest frequencies in all 8 backwaters, especially in Lawrence Lake where 13 species had increased or highest frequencies. The 1998 season seemed to be a continuation of a trend of increased growth begun in previous years, and the status of aquatic vegetation in 1998 was at or near the best since 1991 (Figure 8).

Seventeen species of aquatic plants (13 submersed, 3 floating-leaved and 1 macroalgae) were found in transect sampling in 1998. Coontail (Ceratophyllum demersum), curly, sago, and small or leafy pondweeds, Canadian waterweed, and American white waterlily (Nymphaea odorata) were found in all eight backwaters. The macroalgae chara (Chara spp.) was found in Lawrence Lake for the first time during LTRMP sampling in 1998. Lawrence Lake had the highest species richness (16 species).

Coontail (96%) and Canadian waterweed (68%) were the MFE species in the Stoddard backwater, whereas American white waterlily (91%) and small or leafy pondweeds (88%) were most common in Blue Lake. Coontail decreased from 91% (1997) to 82% (1998) in Blue Lake, dropping it from the first to the third MFE species. Both backwaters are isolated from the main channel. Sago and curly pondweeds and American lotus were the dominant species in the Horseshoe (46%, 24%, and 23%, respectively) and Boomerang (53%, 35%, and 36%, respectively) Island sites, both of which are in the impounded area of Pool 8 and are open to southern winds. Canadian waterweed (56%), coontail (53%), and water stargrass (53%) were the MFE species in Shady Maple. In Lawrence Lake (88%), Goose Island (98%), and Target Lake (55%), coontail was the MFE species. These four backwaters are connected to the main channel. Water stargrass reappeared in three backwaters and increased to highs in all five backwaters in which it was present.

Pool 13

Pool 13 aquatic vegetation, according to data from seven backwaters, had an excellent growing season in 1998 (Figure 9, Table 3). The frequencies of vegetated sites (submersed or rooted floating-leaved aquatic vegetation) and the number of species found either increased from the previous year or remained high in 1998. The frequencies of vegetated sites reached their highest since 1991 in Johnson Creek Levee, Potter's Marsh, Pomme de Terre, and Savanna Bay; increased from the

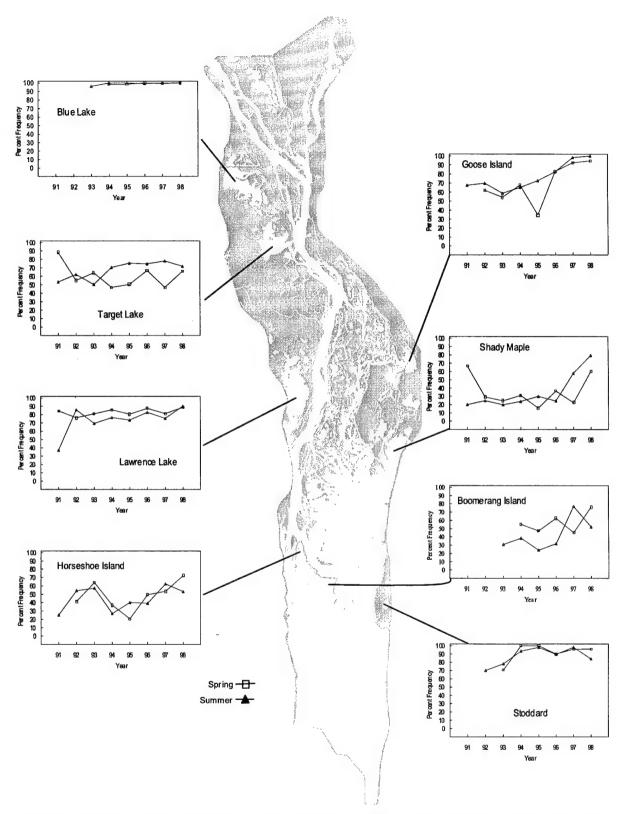
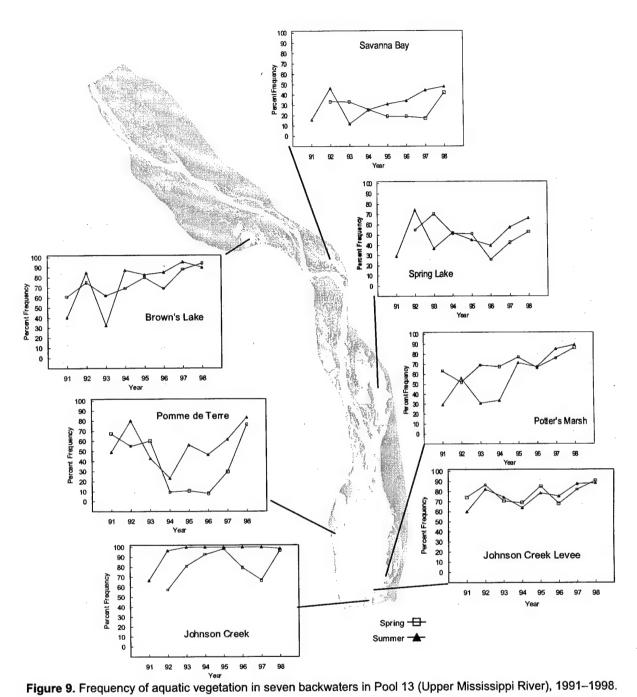


Figure 8. Frequency of aquatic vegetation in eight backwaters in Pool 8 (Upper Mississippi River), 1991–1998.

Table 2. Percent frequency of aquatic plant species by backwater or location in Pool 8, Upper Mississippi River, during spring (Sp. May 19-June 11) and summer (Su, July 21-August 25) sampling periods, 1998.

		1	1	Boomerang	erang	Horseshoe		Shady Manle	Name of the Party	Lawrence	9	Goose Island	Island	Target Lake	Lake	Blue Lake	ake
		Stoddard	Jard	Bland		Signo	1	Silduy	DIA DIA	S	2	S	Su	Sp	Su	Sp	Su
		Sp	DS.	Sp	70	d D	70	de	ne	2	3	3	3		000	10.	1,7,1
Species	u	20	20	<u>5</u>	104	83	81	102	901	434	441	=	115	290	292	134	132
Submersed species											•			,	3 04	;	25 6ab
bladderwort, common	(Utricularia macrorhiza)							,		10.0	13.4	•		r C	0.0	7:11	0.00
buttercup, longbeak	(Ranunculus longirostris)	•		•	•		,				٠,	,		•			
countail	(Ceratophyllum demersum)	96.0 ^b	84.0	1.0		0.9	22.2ªb	36.3	52.8 ²⁶	75.8	87.8ª.b	87.4	98.3**	25.5	54.5	82.1°	81.1
nondweed homed	(Zannichellia palustris)	,		4.84			1.2°										٠,
pondweed curly	(Potamoeton crispus)	56.0°	6.0	34.6 1.		24.1ªb	4.9	6.9	6.0	51.0	21.8	27.0°	13.9	7.6 ^b	5.5	56.9	3.0
pondweed, curry	(P rosteriformis)	•	۷,	•			•	œ œ	34.0ªb	7.7	9.3	53.2	64.3m		•	•	9.0
pondweed, marsient	(D. modomis)			•	۰,			,	6.0	2.1 ^b	6.0	8 .	6:0	0.7	1.4	•	•
pondweed, longleat	(F. nodosus)	22.04	100	52.9b	1.9	45.8b	8.	20.6	7.5	33.1	34.2	11.7	6.1	41.0	33.2	0.9	15.9
pondweed, sago	(1. perinama) (D. foliorus P. pusillus)	0.9	•	1.0		%	6.2	6.9	7.54	32.4ª.b	13.8	21.6	13.9	19.7	21.6b	88.1 ^{a.b}	29.5
pondweeds, small of leafy	(Heterouthern dubia)	; 1	•	0.	27.9ª.b	•	13.6ªb.c	7.8	52.8"b	5.6	10.2ªb	6.3ªb.c	4.3	0.7	3.8 4	•	•
water stangings	(Mericulantallum anicotum)	٠		•	1.0	7.2	17.3	7.8	27.4ªb	38.5	42.6	22.5 ^d	20.0	0.7	3.1		٠
Watermillon, Eurasian	(Myriophyriam spicaram)	ı	•	•		•	12.3ªb		•	3.5	13.6ªb	•	1.7	•	3.8	8.2	17.4
waternymph, nodding	(Najas Jiexilis)	,	, (0	-	13.38	376	55 78.b	×	15.2ªb	51.4	75.7ª.b	4.	10.3	•	3.8
waterweed, Canadian	(Elodea canadensis)	0.89	0.79	6.1	4	7:1	5.71	C.13	1.00	9						1	
wild celery	(Vallisneria americana)	•	•		3.8		4.9°		9.9	0.7	0.5	•	•	•	•	•	•
Rooted floating-leaved species	SS										1		•		9	q2 T	7
American lotus	(Nelumbo lutea)	•	•	28.8	35.6	22.9	17.3	10.8	13.2	14.7	16.3	18.9	9.5	4.77	79.4		
pondlily, yellow	(Nuphar variegata)	2.0	•	•		•		•	•	12.6	9.11	•	٠,	•	. `	7.7	S
waterlily, American white	(Nymphaea odorata)	16.0	16.0	1.0	•	2.4	8.6	2.0	1.9	49.4	51.7	36.9	42.6	23.1	28.4	91.0	90.2
Algae																	
chara	(Chara spp.)	•	•	•	•	•	•	•		•		•	•	•	• !	•	•
filamentous algae	(Chlorophycaea)	48.0	•	1.0	1.0	1.2	4.9	•	5.7	67.6		38.7	42.6		1.7	• ;	• }
Frequency of submersed vegetated sites	etated sites	96.0	84.0	73.1 ^b	31.7	66.3m	44.4	86.9	76.4ª.b	90 90	89.8	91.9	100.04	0.09	67.1	96.3	85.6
Frequency of rooted floating-leaved vegetated sites	-leaved vegetated sites	18.0	6.0	29.8	35.6	25.3	25.9	12.7	14.2	58.5	58.7	52.3	53.9	41.4	49.0	8.16	93.9°
requency of submersed and	Frequency of submersed and/or rooted floating-leaved sites	0.96	84.0	76.0	51.9	72.3 ^b	53.1	8.65	79.2ªb	88.8	90.2ªb	94.6	100.0	629	71.6	0.001	100.0
Significant increase from 1997. PHighest record of 1991–1998. Phispneared between 1997 and 1998.	1908	⁴ Significant decrease from 1997. *Reappeared after absence in 1997. *Lowest record of 1991–1998.	rease from the rease	ce in 1997. 1998.	۲.		<u>F</u>	First record	-ri								
Disappeared over the same			;														

12



previous year in Pomme de Terre and Spring Lake; and remained statistically the same as in 1997 in Johnson Creek and Brown's Lake. Eleven species and one macroalgae reached highs in one or more backwaters. Except for Savanna Bay and Brown's Lake, each of the other five backwaters had at least four taxa that reached highs in 1998. The 1998 growing season appeared to be a continuation of a trend of increase begun in previous years, and the condition of aquatic vegetation in 1998 was at or

near the best since 1991 (Figure 9).

The dominant species varied among the seven backwaters. In Johnson Creek, the southernmost study area in Pool 13, American lotus was the MFE species (91%), followed closely by water stargrass and wild celery both at 88%. Wild celery was the MFE species on Johnson Creek Levee at 85% followed by water stargrass at 76%. The reverse was apparent on Potter's Marsh where water stargrass was present at 80% of the sites and wild celery 73% of the sites. On the basis of visual inspection during the sampling, we estimated that

Table 3. Percent frequency of aquatic plant species by backwater or location in Pool 13, Upper Mississippi River, during spring (Sp. May 21–June 23) and summer (Su; August 3–September 3) sampling periods, 1998.

Species Submersed species Contail pondweed, horned (Zannichellia palustris) pondweed, curly pondweed, flatstem (P. zosteriformis) pondweed, longleaf (P. nodosus) pondweed, sago (P. pectinatus) pondweed, sall or leafy (P. foliosus, P. pusillus) water stargrass waternymph, nodding (Najas flexilis) waterweed, Canadian (Kyriophyllum spicatum) waterweed, Canadian (Elodea canadensis) wild celery (Vallisneria americana) Roched floatine-leaved species*	Sp Su Su 66 67 67 66 67 67 66 67 67 68 67 67 68 68 68 68 68 68 68 68 68 68 68 68 68		2		Sp Su 110 132 4.5 16.7tb - - 3.6 - 2.7b 1.5 54.5 9.1 - - 2.7b 1.5 54.5 9.1 - - 25.5 80.3b 51.8tb 47.0 0.9 4.5	38.2 68 68 68 68 68 7 7 1.5 7 22.1		Sp Su 179 153 179 179 170 172 172 172 172 172 173 174 174 174 174 174 175	35.2* 35.2* 	Sp Su 144 143 28.5 39.9 th 		Sp Su 432 422 68.8 84.8*b 0.5° - 14.6* - 3.0 0.7 56.5′ 28.9	Su 422 422 0.7 28.9 0.5
(Ceratophyllum demersum) (Cannichellia palustris) (Potamogeton crispus) (P. costeriformis) (P. pectinatus) I or leafy (P. foliosus, P. pusillus) (Heteranthera dubia) (Heteranthera dubia) (Ayriophyllum spicatum) Iding (Najas flexilis) (Hodea canadensis) (Vallisneria americana)					Su 132 16.7 th - - 1.5 9.1 - 80.3 ^b 47.0	Sp 68 68 78.2	85 61.2* - 2.4 4.7 8.2 7.1*** 70.6**	153 13.7 7.2 7.2 7.8 34.0° 5.9*b	35.2* 35.2* 6.7 6.7 0.6	28.5 28.5 - 1.4 1.4 - 18.1 0.7°	39.9 ⁴ b		84.8 ^{ab} 84.8 ^{ab} 0.7 0.7 0.5
(Ceratophyllum demersum) d (Zannichellia palustris) (Potamogeton crispus) af (P. zosteriformis) af (P. pectinatus) (P. pectinatus) (P. pectinatus) (P. pectinatus) (Heteranthera dubia) asian (Myriophyllum spicatum) dding (Najas flexilis) tthem (N. guadalupensis) udian (Elodea canadensis) (Vallisneria americana)			4.6		132 16.7 th - 1.5 9.1 - 80.3 ^b 47.0	38.2	85 61.2* 2.4 4.7 8.2 7.1abs 70.6*b	153 13.7 7.2 7.8 34.0° 5.9**	35.2*	28.5 - 1.4 - - - - 18.1 0.7°	39.9*b 0.7 7.7 - 12.6*b		84.8*b 0.7 0.7 28.9
(Ceratophyllum demersum) (Cannichellia palustris) (Potamogeton crispus) (P. tosteriformis) (P. tosteriformis) (P. pectinatus) (P. pectinatus) (P. petinatus) (Heteranthera dubia) (Heteranthera dubia) (Hojas flexilis)					16.74b	38.2 - 4.4 - 4.4 35.3* - 1.5	61.2*	13.7 - 7.2 - 7.8 34.0° 5.9*b	35.2*	28.5 - 1.4 - - 18.1 0.7° 4.9	39.94b - 0.7 - 7.7 - 12.64b		84.8*b - - 0.7 28.9
(Ceratophyllum demersum) (Zannichellia palustris) (Potamogeton crispus) (P. zosteriformis) (P. nodosus) (P. pectinatus) (P. foliosus, P. pusillus) (Heteranthera dubia) (Heteranthera dubia) (Myriophyllum spicatum) (Najas flexilis) (N. guadalupensis) (Elodea canadensis)					16.7# 1.5 9.1 - 80.3* 47.0	38.2 . 4.4 . 4.4 35.3* 1.5	61.2* - 2.4 - 4.7 8.2 7.1*** 70.6**	7.2 7.2 7.8 34.0° 5.9 ^{ab}	33.2. 		0.7 0.7 7.7 12.6*b		0.7 0.7 0.8 0.5
(Zannichellia palustris) (Potamogeton crispus) (P. zosteriformis) (P. nodosus) (P. pectinatus) (P. foliosus, P. pusillus) (P. foliosus, P. pusillus) (Ayriophyllum spicatum) (Najas flexilis) (N. guadalupensis) (Elodea canadensis) (Vallisneria americana)					- 1.5 9.1 - 80.3 47.0	4.4 4.4 35.3* - 1.5 22.1	2.4 - 4.7 8.2 7.1 abs 70.6 ab	7.2 7.8 34.0° 5.9*b	8.4 6.7 0.6 29.1*b	1.4 1.4 1.8.1 0.7°	0.7 - 7.7 -	9.5° 14.6° 3.0 56.5° 6.3°b	0.7
(Potamogeton crispus) (P. zosteriformis) (P. nodosus) (P. pectinatus) (P. foliosus, P. pusillus) (Heteranthera dubia) (Myriophyllum spicatum) (Najas flexilis) (N. guadalupensis) (Elodea canadensis) (Vallisneria americana)					1.5 9.1 80.3 47.0	4.4 4.4 35.3* - 1.5 22.1	2.4 4.7 8.2 7.1 ^{abs} 70.6 ^{ab}	7.2 7.8 34.0° 5.9*b	8.4 6.7 0.6 29.1*b	1.4 - - 18.1 0.7° 4.9	0.7 - 7.7 -	14.6* - 3.0 56.5 ^f 6.3*b	- 0.7 28.9 0.5
(P. zosteriformis) (P. nodosus) (P. pectinatus) (P. foliosus, P. pusillus) (Heteranthera dubia) (Myriophyllum spicatum) (Najas flexilis) (N. guadalupensis) (Elodea canadensis) (Vallisneria americana)					1.5 9.1 - 80.3 ^b 47.0	4.4 35.3* - 1.5 22.1	- 4.7 8.2 7.1ªb£ 70.6ªb 29.4	7.8 34.0° 5.94b 13.1	8.4 6.7 0.6 29.1*b	18.1 0.7° 4.9	7.7	3.0 56.5 ^f 6.3*b	- 0.7 28.9 0.5
(P. nodosus) (P. pectinatus) (P. foliosus, P. pusillus) (Heteranthera dubia) (Myriophyllum spicatum) (Najas flexilis) (N. guadalupensis) (Elodea canadensis) (Vallisneria americana)					1.5 9.1 - 80.3 47.0 4.5	4.4 35.3° - 1.5 22.1	4.7 8.2 7.1*b# 70.6*b 29.4	7.8 34.0° 5.94b 13.1	8.4 6.7 0.6 29.1 ^{ab}	18.1 0.7° 4.9	7.7 12.6*b	3.0 56.5 ^f 6.3 ^{ab}	0.7 28.9 0.5
(P. pectinatus) (P. foliosus, P. pusillus) (Heteranthera dubia) (Myriophyllum spicatum) (Najas flexilis) (N. guadalupensis) (Elodea canadensis) (Vallisneria americana)					9.1 80.3 ^b 47.0 4.5	35.3* - 1.5 22.1	8.2 7.1 ^{ab.g} 70.6 ^{ab} 29.4	34.0° 5.9*b 13.1	6.7 0.6 29.1*b	18.1 0.7° 4.9	7.7 - 12.6*b	56.5 ^t 6.3 ^{±b} -	28.9
(P. foliosus, P. pusillus) (Heteranthera dubia) (Myriophyllum spicatum) (Najas flexilis) (N. guadalupensis) (Elodea canadensis) (Vallisneria americana)					80.3° 47.0 4.5	- 1.5 22.1	7.1°bs 70.6°b 29.4	5.94b 13.1	0.6 29.1 ^{ab}	0.7°	12.6 ^{a,b}	6.3*b	0.5
(Heteranthera dubia) (Myriophyllum spicatum) (Najas flexilis) (N. guadalupensis) (Elodea canadensis) (Vallisneria americana)					80.3 ^b 47.0 4.5	1.5	70.6*b 29.4	13.1	29.146	4.9	12.6ªb		
(Myriophyllum spicatum) (Najas flexilis) (N. guadalupensis) (Elodea canadensis) (Vallisneria americana)				51.8 ² b 0.9	47.0	22.1	29.4	:					•
(Najas flexilis) (N. guadalupensis) (Elodea canadensis) (Vallisneria americana)				0.0	4.5			=	12.8ªb		0.7	•	
(N. guadalupensis) (Elodea canadensis) (Vallisneria americana)	. 29.			-			4.7	9.2	33.0ab				•
(Flodea canadensis) (Vallisneria americana)			0.7	×.	1.5		7.16	•	P.,	•			•
(Vallisneria americana)	1.5	. 6.71	. 13.2	1.8	29.5*	11.8	45.9ªb	4.6	10.1ªb	,	0.7	0.2 °	•
ooted floating-leaved species*	66.7 88		81.5 85.1 ^{ab}	•	72.7ªb	39.7	49.4mp	0.7	4.5ªb	•	,		•
(Not beautiful freeze)	40.0	0 ا ق	219	2.7	6.1	7.4	10.6	18.3	50.8*b	8.3	21.7	85.6	0.09
(paint nominal)						0	106	2.0	3.4			0.2	0.2
waterlily, American white (Nymphaea odorata)	12.1	37.3	5.6 14.9			0.0	0.01	2	Š				
Algae*													
chara (Chara spp.)	. 7.5	7.5ab.s 0	- 6.0	•		•	•	0.7	•	•	•	. ;	• ;
filamentous algae (Chlorophycaea) 6	60.6 2.	23.9 30	36.1 69.3 ^{a,b}	ъ 29.1	53.8	13.2	48.2ªb	26.1**	7.8	8.3	17.5**	55.1	17.3
nitella (Nitella spp.)		1.5°	•	•		•	•	•		•			•
Frequency of submersed vegetated sites	93.9	98.5 90	90.7 88.6	86.4	89.4 ^b	75.0	83.5ªb	52.3	53.1	38.9	45.5 ^{a,b}	80.1	85.8
etated sites	50.0	97.0	13.9 21.9	2.7	6.1	16.2	21.2	19.0	\$0.8	8.3	21.7	85.6	0.09
ed sites	97.0	98.5 90	90.7 88.6	86.4	89.4	76.5	83.5*b	52.9	66.5	42.4	48.3°	94.0	90.0
Significant increase from 1997 and 1998	red betwe	en 1997	and 1998.			⁸ Reapp	Reappeared after absence in 1997	er absenc	e in 199	7.			

 American lotus would have had higher than recorded frequencies in Potter's Marsh and Johnson Creek Levee if transects in the two backwaters were extended to the shoreline. Because the transects were established in 1991 and their locations remain unchanged, a large portion of the American lotus beds between the shoreline and the starting points of transect were missed in the transect sampling. In Pomme de Terre, the northernmost impounded area sampled, water stargrass was the MFE species (71%), followed by coontail (61%). American lotus was the MFE species in Spring Lake (51%) and Brown's Lake (86%), whereas coontail was a MFE species (40%) in Savanna Bay.

Wild celery and water stargrass reached highs for the period 1991-98 in five backwaters and increased from 1997 in four backwaters. Water stargrass was present in six backwaters whereas wild celery was present in five. Small or leafy pondweeds were found in Johnson Creek Levee and Savanna Bay for the first time and reappeared in Pomme de Terre at its highest frequency for the period 1991-98. Horned pondweed was found in Brown's Lake for the first time but disappeared from Johnson Creek. Flatstem pondweed disappeared from Johnson Creek Levee; and southern waternymph (Najas guadalupensis) disappeared from Spring Lake. Eurasian watermilfoil, an invasive exotic species, was found in six of the seven backwaters and reached highest frequencies in Potter's Marsh and Spring Lake for the period 1991-98. It reappeared in Savanna Bay after an absence in 1997, and its frequency in Spring Lake doubled from 1997 to 1998.

Seventeen species (not including filamentous algae) were found in Pool 13 in 1998. Chara, nitella (Nitella spp.), small or leafy pondweeds, Canadian waterweed, Eurasian watermilfoil, and horned pondweed reappeared after an absence in 1997 or made a first appearance in one or more locations. Flatstem and horned pondweeds were unique to the spring sampling episode and nitella was unique to summer. Coontail, curly and sago pondweed, Canadian waterweed, and American lotus were found in all seven backwaters.

According to notes of casual observations, some Eurasian watermilfoil leaflets on the Johnson Creek Levee transects appeared to have been eaten this year, possibly by insect or fish activity. A species of boring worm was observed in large numbers on American lotus in Brown's and Spring Lakes for the last several years. Large numbers of a katydid were observed on the American lotus in Spring Lake this year. Some of the leaves were partially striped.

Pool 26

In general, aquatic vegetation has been rare in contiguous backwater areas in Pool 26 in recent years, especially since a devastating flood in 1993, and based on data from four backwaters, we conclude that the status of aquatic vegetation remained unchanged in 1998. Aquatic vegetation was found in all four backwaters surveyed in 1998, three of the four had no aquatic vegetation in 1997 (Figure 10, Table 4). Stump Lake had one of the highest frequencies of aquatic vegetation since 1991, whereas the other three backwaters remained virtually nonvegetated. Sago pondweed was the only species found in Calhoun Point (1%) and Swan Lake (3%). Coontail (3%) and wingleaf primerosewillow (Ludwigia decurrens; 7%) were the only species found in Fuller Lake. Stump Lake is an isolated backwater managed as a moist soil unit. The lake is separated from the main channel of the Illinois River by a low, constructed levee system and a rather wide riparian forest buffer. Sago pondweed was the MFE submersed aquatic plant in Stump Lake during 1998 (92%) and exhibited a significant increase in frequency from 1997. Coontail and longleaf pondweed (Potamogeton nodosus) were the next most abundant (40%). These species also exhibited significant increases in frequency over their 1997 status. American lotus exhibited a significant increase in frequency between 1997 and 1998 levels. American lotus is most dominant in the lower unit of Stump Lake. Wingleaf primerosewillow reappeared in 1998 along transects in Stump Lake following an absence in 1997 and was most often found growing as thick mats along the margins of the lake.

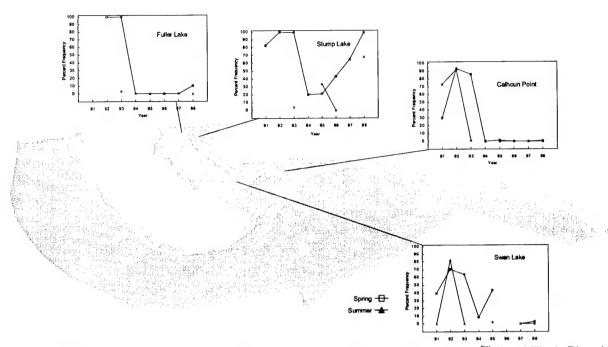


Figure 10. Frequency of aquatic vegetation in four backwaters in Pool 26 (Upper Missisšīppi and Illinois Rivers), 1991–1998.

Table 4. Percent frequency of aquatic plant species by backwater or location in Pool 26, Upper Mississippi and Illinois Rivers, during spring (Sp; May 18–June 3) and summer (Su; August 4–August 11) sampling periods, 1998.

		Calhou	n Point	Swan	Lake	Stump	Lake	Fuller	Lake
		Sp	Su	Sp	Su	Sp	Su	Sp	Su
Species	n	150	157	267	292	169	175	29	29
Submersed species									
coontail	(Ceratophyllum demersum)	-	-	-	-	39.6ª	32.8	3.4 ^b	-
pondweed, longleaf	(Potamogeton nodosus)	-	-	-	-	36.1	39.7 ^{a,c}	-	-
pondweed, sago	(P. pectinatus)	0.7^{b}	-	$2.6^{a,b}$	-	92.3 ^{a,c}	25.9	-	-
pondweeds, small or leafy	(P. foliosus, P. pusillus)	-	-	_	-	0.6	3.4 ^{a,b}	-	-
water stargrass	(Heteranthera dubia)	-	-	-	-	-	0.6^{b}	-	-
waterweed, Canadian	(Elodea canadensis)	-	-	-	-	0.6^d	0.6	-	-
Rooted floating-leaved spec	ies								
American lotus	(Nelumbo lutea)	_	-	-	-	33.1ª	27.6	-	-
pondlily, yellow	(Nuphar variegata)	-	-	-	_	_	_ c	-	-
primrosewillow, wingleaf	•	-	-	-	-	5.3 ^{a,b}	2.9	6.9 ^b	-
Algae									
filamentous algae	(Chlorophycaea)	_	_	_	_	-	1.7 ^{a,b}	-	_
Frequency of submersed site		0.7 ^b	_	2.6 ^{a,b}	-	98.8ª	58.0	3.4 ^b	-
Frequency of rooted floating		_	_	_	_	35.5ª	29.3	6.9 ^b	_
Frequency of submersed or		0.7 ^b	-	2.6 ^{a,b}	-	99.4ª	67.8	10.3 ^{a,b}	-

^aSignificant increase from 1997.

^cHighest record of 1991–1998.

^eDisappeared between 1997 and 1998.

^bReappeared after absence in 1997.

^dSignificant decrease from 1997.

Frequency of submersed aquatic vegetation among transect sites in Pool 26 backwaters continued to exhibit two distinct patterns related to the 1993 flood. Before 1993, all four backwaters sampled using transect methods contained substantial amounts of submersed aquatic vegetation. Immediately following the 1993 flood, transect surveys indicated the plant communities were nearly eliminated. One pattern evident during the years following the flood can be described as a "recovery." The aquatic plant community in Stump Lake today is very similar to its pre-flood status. This recovery took 5 years. The second pattern can be described as a "crash without recovery." Fuller Lake, Calhoun Point, and Swan Lake continue to support little to no submersed aquatic vegetation.

The environmental conditions favoring or preventing plant community recovery within each of the four backwater lakes sampled along Pool 26 have not been identified. However, observations indicate that Stump Lake may have physical conditions that have favored plant reestablishment following the 1993 flood. For example, its location is farther away from the main river channel than the other backwaters sampled. Between the lake and river are floodplain forests that may act to slow down floodwaters and allow sediments to drop out of suspension before they reach Stump Lake. This improved water clarity would encourage reestablishment of submersed aquatic vegetation. As plants slowly became established, they too would tend to improve water clarity. Conversely, the backwaters that do not support submersed aquatic vegetation are much closer to the main river channel and the water is often more turbid throughout the growing season.

La Grange Pool

Aquatic vegetation has been rare in contiguous backwater areas in La Grange Pool in recent years, especially since a devastating flood in 1993, and based on data from four backwaters, the status of aquatic vegetation remained unchanged in 1998. Aquatic vegetation was found in three of the four backwaters surveyed in 1998; one of the three had no aquatic vegetation in 1997 (Figure 11, Table 5). Grape Island is a side channel border where sago

pondweed was found in 1991-94 but has not had aquatic vegetation present during the sampling windows since 1996. However, sago and horned pondweeds were observed in Grape Island during 1995-98 after the summer sampling window closed, often during late September and early October. Unstable water levels within La Grange Pool seemed to have been a factor in this late season growth with the more stable water conditions allowing the plants to establish late in the growing season. Although aquatic vegetation was found during the transect sampling in Banner Marsh and Spring Lake, a stratified random sampling conducted between June 15 and July 15 revealed no submersed aquatic vegetation at any of the 550 sites within La Grange Pool.

Point Lake is a semi-isolated backwater where several aquatic plant species have been found since 1992 including coontail, sago pondweed, and western waterweed (*Elodea nuttali*). Following the flood of 1993, aquatic vegetation within Point Lake has steadily declined until disappearing completely in 1997. Coontail (44%) and wingleaf primrosewillow (32%) reappeared in Point Lake in 1998. Before 1996, coontail was the most abundant species in Point Lake, present at a percent frequency as high as 80% in 1993.

In Spring Lake, Eurasian watermilfoil has been present since the late 1980s. Percent frequencies have been as high as 93% in 1997 and as low as 32% following an out break of aquatic milfoil weevil coupled with an unusually high surface temperature (>108° F) during the summer of 1995 (Wayne Herndon, Iowa Department of Natural Resources fisheries, personal communication). In 1998, percent frequencies of Eurasian watermilfoil, the MFE species, had a significant drop to 63%. Flatstem pondweed, water stargrass, brittle waternymph (Najas minor), and Canadian waterweed disappeared along transects in 1998. Although low in abundance (1%), longleaf pondweed made a reappearance after being absent in 1997. Percent frequency of American lotus was recorded at a high of 14% in 1998. Over all percent frequency of sites with submersed aquatic plants

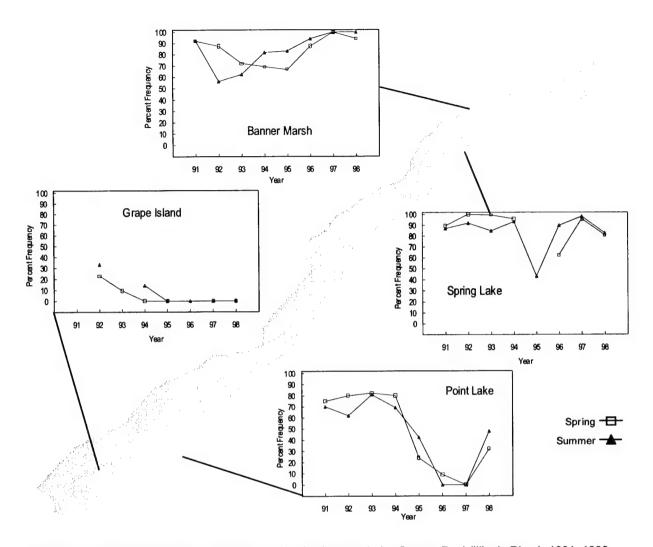


Figure 11. Frequency of aquatic vegetation in four backwaters in La Grange Pool (Illinois River), 1991–1998.

along transects at Spring Lake were down significantly from 99% in 1997 to 83% in 1998.

In Banner Marsh, the frequency of coontail, the MFE species, remained statistically the same as in 1997. Longleaf pondweed increased over the previous year, small or leafy pondweeds reappeared, and water grass and American lotus were first found in 1998. The presence of American lotus in Banner Marsh could be related in part to the connection of Bulrush Pond to an adjacent lotus-filled lake after water elevations in Banner Marsh were raised through the HREP project there.

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Table 5. Percent frequency of aquatic plant species by backwater or location in La Grange Pool, Illinois River, during spring (Sp, May 22-June 20) and summer (Su, July 29-August 11) sampling periods, 1998.

0 1 0 (1)		Grape	Island	Point I	Lake	Spring	Lake	Banner	Marsh
		Sp	Su	Sp	Su	Sp	Su	Sp	Su
Species	n	16	18	25	23	67	86	18	18
Submersed species									
coontail	(Ceratophyllum demersum)	-	-	-	43.5 ^{a,b}	4.5	8.1°	72.2	83.3
pondweed, horned	(Zannichellia palustris)	-	-	-	-	7.5	-	-	-
pondweed, curly	(Potamogeton crispus)	-	-	-	-	10.4 ^d	-	11.1 ^b	5.6
pondweed, flatstem	(P. zosteriformis)	-	-	-	-	-	_d,e	-	-
pondweed, longleaf	(P. nodosus)	-	-	-	- '	-	1.2 ^b	16.7 ^b	33.3ª
pondweed, sago	(P. pectinatus)	-	-	-	-	6.0	5.8	33.3^{f}	22.2
pondweeds, small or leafy	(P. foliosus, P. pusillus)	-	-	-	-	4.5	-	-	5. 6 ^b
water stargrass	(Heteranthera dubia)	٠.	-	-	-	-	- e	-	11.1 ^g
watermilfoil, Eurasian	(Myriophyllum spicatum)	_	-	-	• .	61.2	62.8 ^d	50.0	61.1
waternymph, brittle	(Najas minor)	-	-	-	-		_ e	-	-
waternymph, nodding	(N. flexilis)	-	-	-	-	1.5	4.7	-	-
waterweed, Canadian	(Elodea canadensis)	-	-	-	-	-	_d,e	-	-
Rooted floating-leaved spec	ies								
American lotus	(Nelumbo lutea)	-	-	-	-	11.9	14.0 ^f	-	5.6 ^g
primrosewillow, wingleaf	(Ludwigia decurrens)	-	-	32.0 ^{a,b,f}	13.0	-	-	-	-
waterlily, American white	(Nymphaea odorata)	-	-	-	-	4.5	4.7	-	-
Algae									
chara	(Chara spp.)	_	-	-	-	10.4	10.5	-	-
filamentous algae	(Chlorophycaea)	-	-	-	-	16.4 ^d	8.1	16.7 ^d	11.1
Frequency of submersed site	es	_	_	_	43.5 ^{a,b}	76.1	82.6 ^d	94.4	100.0
Frequency of rooted floating		_	_	32.0 ^{a,b,f}		16.4	18.6	_	5.6 ^g
Frequency of submersed or		_	-	32.0	47.8 ^{a,b}	80.6	82.6 ^d	94.4	100.0

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gFirst record.

^aSignificant increase from 1997. ^bReappeared after absence in 1997.

^cLowest record of 1991–1998.

^dSignificant decrease from 1997.

^eDisappeared.

Highest record of 1991-1998.

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Locations, Habitat, Number of Transects and Sites, Sampling Dates, and Distances Between Sites Sampled in Pools 4, 8, 13, and 26 of the Mississippi River and La Grange Pool of the Illinois River During the 1998 Sampling Season.

Appendix A

		Number of transects	Number of sites	Number of sites	Sampling dates	Sampling dates	Distance between sites (m)
Location	Habitat	spring:summer	spring	summer	spring	summer	sites (m
Pool 4							
Mud Lake (M791.3) ^a	BWC	3:3	57	54	6/17	8/27, 9/2	30
Dead Slough Lake (M789.2, M788.5, M788.0)	BWC	9:9	134	133	6/11 B 12, 16	8/19, 27, 9/2	30
Goose Lake (M788.G) ^b	BWC	3:3	25	24	6/17	9/2	30
Catherine Pass (Bay City Flats ^c ; M787.0)	BWC	3:3	71	74	6/16	8/27, 9/2	30
Robinson Lake (M758.R) ^b	BWC	9:9	203	218	6/1, 3 B 5, 10	8/18, 20, 24B25	30
Rice Lake/Big Lake Bay (M758.0, M758.5 ^d)	BWC	6:6	71	67	6/8, 10	8/26	30
Big Lake (M757.5) ^d	BWC	5:5	157	148	6/8, 10	8/21, 25 B 26	30
Upper Peterson Lake (M754.8, M754.5)	BWC	6:6	69	97	5/26828	8/12B13	30
Lower Peterson Lake (M753.5)	BWC	4:3	130	86	5/28B29	8/17	30
Total		48:47	917	901	14	12	
Pool 8							
Blue Lake (M697.0)	BWI	3:3	134	132	5/26	8/10	15
Target Lake (M696.0)	BWC	11:11	290	292	5/19, 21	7/21, 29B31	15
Goose Island (M692.0)	BWC	5:5	111	115	6/2	8/11	15
Lawrence Lake (M691.0)	BWC	10:10	434	441	6/4B5, 8, 10Bl 1	8/13 B 14, 19, 21, 24 B 25	15
Shady Maple (M690.0)	BWC	3:3	102	106	5/29	8/12	15
Horseshoe Island (M687.0) ^d	BWC	5:5	83	81	6/3	8/7	15
Boomerang Island (M686.0) ^d	IMP	4:4	104	104	5/22	8/3	15
Stoddard (M684.0)	BWI	4:4	50	50	5/27	8/4	15
Total:		45:45	1308	1321	13	16	

Location	Habitat	Number of transects spring:summer	Number of sites spring	Number of sites summer	Sampling dates spring	Sampling dates summer	Distance between sites (m)
	Habitat						
Pool 13 Brown's Lake (M545.1, M544.5) ^d	BWC	20:20	432	422	6/4B5, 8, 10Bl 1, 17Bl 9, 22B23	8/14, 18B21, 24B26	15
Savanna Bay (M541.5, M540.5, M539.5)	BWC	12:12	144	143	5/26B27	8/6B7, 11	15
Spring Lake (M534.8, M533.6, M532.0)	BWC	12:12	153	179	6/1B2, 4	8/11B14	15
Pomme de Terre (M526.0)	IMP	5:5	68	85	5/21	8/3	15
Potter's Marsh (M524.0) ^d	IMP	6:6	110	132	6/12, 15	8/27, 31	15
Johnson Creek Levee (M523.5)	IMP	4:4	108	114	6/16, 19	9/2B3	15
Johnson Creek (M523.0)	IMP	2:2	66	67	5/22, 26	8/5	15
Total:		61:61	1081	1142	19	19	
Pool 26							
Calhoun Point (I003.0) ^e	BWI	17:18	150	157	5/21, 26	8/11	15
Swan Lake (I005.5) ^d	BWI	10:10	267	292	6/2, 3	8/4	15
Stump Lake (I010.0)	BWI	8:8	169	175	5/18B19	8/6, 10	15
Fuller Lake (1011.5)	BWI	2:2	29	29	6/2	8/6	15
Total:		37:38	615	653	6	4	
LaGrange Pool							
Grape Island (I086.4)	МСВ	3:3	16	18	6/10	8/11	15
Point Lake (I100.0)	BWI	6:6	25	23	5/22	8/4	15
Spring Lake (1135.5)	BWI	5:5	67	86	6/4, 18, 20	7/29, 31, 8/3	15
Banner Marsh (Bulrush Pond ^e ; I140.0) ^d	BWI	2:2	18	18	6/5	8/7	15
Total:		16:16	126	145	6	6	

^a Mississippi River miles, measured from the confluence of the Mississippi and Ohio Rivers.

^b AG≡and AR≡to distinguish this lake from another lake with the same river mile.

^c Locally recognized alternate name.

^d Part of a Habitat Rehabilitation and Enhancement Project.

^e Pool 26 is located at the confluence of the Illinois and Mississippi Rivers and the portions named here extend up the Illinois River, are managed by the Illinois Department of Natural Resources, and are designated by Illinois River miles. Illinois River miles are measured from the confluence of the Illinois and Mississippi Rivers. measured from the confluence of the Illinois and Mississippi Rivers.

Appendix B

List of Submersed and Floating-leaved Species Found During LTRMP Monitoring in Pools 4, 8, 13, and 26 of the Mississippi River and La Grange Pool of the Illinois River, 1991–1998.ª

Family	Code	Scientific name	Common name
Azollaceae	AZME, AZCA	Azolla spp.b	watervelvet, mosquitofern
Ceratophyllaceae	CEDE4	Ceratophyllum demersum L.	coontail, coon=s tail
Characeae	CHAR	Chara spp.	chara
Characeae	NITE	Nitella spp.	nitella
Haloragaceae	MYSI	Myriophyllum sibiricum Komarov	northern watermilfoil, short spike watermilfoil
Haloragaceae	MYSP2	M. spicatum L.	Eurasian watermilfoil, spike watermilfoil
Hydrocharitaceae	ELCA7	Elodea canadensis Michx.	Canadian waterweed
Hydrocharitaceae	ELNU2	E. nuttallii (Planch.) St. John c	western waterweed
Hydrocharitaceae	VAAM3	Vallisneria americana Michx.	wild celery, American eelgrass
Lemnaceae	LEMI3	Lemna minor L.b	lesser duckweed, small duckweed, common duckweed
Lemnaceae	LETR	L. trisulca L.b	star duckweed
Lemnaceae	SPPO	Spirodela polyrhiza (L.) Schleid. ^b	greater duckweed, big duckweed, common duckweed
Lemnaceae	WOBR	Wolffia braziliensis Weddell ^b synonymy W. papulifera C. Thompson and W. punctata Griseb.	Brazillian watermeal
Lemnaceae	woco	W. columbiana Karst.b	Columbian watermeal
Lentibulariaceae	UTMA	Utricularia macrorhiza Le Conte, synomny Utricularia vulgaris L.	common bladderwort
Najadaceae	NAFL	Najas flexilis (Willd.) Rostkov and Schmidt	bushy pondweed, slender naia nodding waternymph
Najadaceae	NAGR	N. gracillima (A. Braun ex Engelm.) Magnus	slender waternymph
Najadaceae	NAGU	N. guadalupensis (Spreng.) Magnus ^d	southern waternymph
Najadaceae	NAMI	N. minor All.	brittle waternymph
Nymphaeaceae	NELU	Nelumbo lutea Willd.	American lotus
Nymphaeaceae	NULU	Nuphar variegata Durand synonymy N. lutea (L.) Sm.°	yellow pondlily
Nymphaeaceae	NYOD	Nymphaea odorata Ait. synonymy N. tuberosa Paine	American white waterlily
Onagraceae	LUDE4	Ludwigia decurrens Walt.c	wingleaf primrosewillow
Onagraceae	LUDE4	Ludwigia decurrens Walt.	wingleaf primrosewillow

Family	Code	Scientific name	Common name
Pontederiaceae	ZODU	Heteranthera dubia (Jacq.) MacM. synonymy Zosterella dubia (Jacq.) Small	water stargrass, grassleaf mudplantain
Potamogetonaceae	POAL8	Potamogeton alpinus Balbis ^c	red pondweed, alpine pondweed
Potamogetonaceae	POCR3	P. crispus L.	curly pondweed, curlyleaf pondweed
Potamogetonaceae	POEP2	P. epihydrus Raf.	ribbonleaf pondweed
Potamogetonaceae	POFO3	P. foliosus Raf.	leafy pondweed
Potamogetonaceae	POGR8	P. gramineus L.	variableleaf pondweed
Potamogetonaceae	POIL	P. illinoensis Morong.	Illinois pondweed
Potamogetonaceae	PONO2	P. nodosus Poir	river pondweed, American pondweed, longleaf pondweed
Potamogetonaceae	POPE6	P. pectinatus L	sago pondweed
Potamogetonaceae	POPU7	P. pusillus L.	small pondweed, slender pondweed
Potamogetonaceae	PORI2	P. richardsonii (Benn.) Rydb.	Richardson-s pondweed
Potamogetonaceae	POZO	P. zosteriformis Fern.	flatstem pondweed
Ranunculaceae	RAFL	Ranunculus flabellaris Raf.	yellow water buttercup
Ranunculaceae	RALO2	R. longirostris Godr. f	longbeak buttercup
Zannichelliaceae	ZAPA	Zannichellia palustris L.	horned pondweed

^a Scientific nomenclature and common names follow the U.S. Department of Agriculture Internet PLANTS Database

^{(1996).} Common names used by Upper Mississippi River managers are also included.

^b Species excluded from analysis.

^c Species verified by Dr. C.B. Hellquist, North Adams State College, Massachusetts.

^d Species verified for Pool 13 by Dr. E. Cawley, Loras College, Iowa, and for Pool 8 by Dr. C. B. Hellquist, North Adams State College, Massachusetts.

Scientific nomenclature follows Gleason and Cronquist (1991). Nuphar lutea spp. variegata in PLANTS database.

Ranunculus longirostris and R. trichophyllus were combined (Voss 1985).

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² Wisconsin Department of Natural Resour ³ Illinois Natural History Survey, 8450 Mo ⁴ Iowa Department of Natural Resources, 2 ⁵ Illinois Natural History Survey, 704 Nort	Environmental Sciences Center, 2630 Fanta R ces, 575 Lester Avenue, Onalaska, Wisconsir ntclair Avenue, Brighton, Illinois 62012	1 54650	, 54603	8. PERFORMING ORGANIZATION REPORT NUMBER
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of changes of submersed and floatin permanent transects. Species compos calculated and compared with results upper Pool 4 (above Lake Pepin), wh aquatic vegetation declined from 1991 13 since 1994 and is at or near its best vegetation has been rare in La Grange	g-leaved aquatic vegetation. Aquatic ve itions, frequencies of individual species, is from previous years. The status and treere aquatic vegetation has declined since to 1996 and started to rebound in 1997, t (recorded) condition since 1991. In recer Pool (Illinois River) and Pool 26. In Pool vegetation in 1997. Similarly, aquatic veg	getation was sampled at regu- and the frequencies of sites and of aquatic vegetation in 1- 1991, the trend continued in 1- he rebound continued in 1998, the years, especially since a dev- 26, however, aquatic vegetati	ilarly sp that sup 998 vari 1998. In Aquation astating on was	in 1998 to determine the status and trend laced sites along previously established exported aquatic vegetation in 1998 were ded among the thirty-two backwaters. In lower Pool 4 (below Lake Pepin), where c vegetation has increased in Pools 8 and flood in 1993 and through 1998, aquatic found in all four backwaters surveyed in backwaters surveyed in La Grange Pool
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